

Application of Vermicompost, Spent Mushroom Substrate, Domestic Compost and Leachate as Inoculum on Bioremediation of Oil Sludge

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Abstract

The research on utilization of spent mushroom substrate (SMS), vermicompost (VC), domestic compost (DC) and leachate (L) as inoculum of microorganisms in bioremediation of 30% oil sludge was conducted. The aim of this study was to obtain the most suitable inoculum in bioremediation of oil sludge. TPH and C/N ratios were analyzed on 15 days of interval, while Polycyclic Aromatic Hydrocarbon (PAH) and heavy metals were measured at the beginning (H-0) and the end of experiment (H-45) for 45 days of experiment. The additional parameters such as acidity (pH), temperature (°C), and humidity of medium (%) were measured. The results showed that all inoculums (SMS, VC, DC and L) affect the diminution of TPH as much as 27.7%, 20.9%, 24.8% and 24.20%, respectively. On the other hand, SMS and VC were able to decrease C/N ratio as much as 18% and 4%, respectively. By contrast, DC and L increased C/N ratio as much as 33% and 1.14%, respectively. In this study, inoculum of SMS and L were able to decrease Cr content (16% and 5.88%, respectively). However, Hg content diminished (29%) only in oil sludge medium with SMS inoculum. The PAH compound can be degraded by all types of inoculum with carbon chain ranged from C₆-C₃₄ into shorter carbon chain (C₆-C₃₀).

Keywords: Bioremediation; Oil sludge; Biofertilizer

Introduction

Petroleum waste is considered as industrial waste from the oil exploration and refining processes. This waste potentially produced oil sludge that contains oil, water, ash, rust, sand and other materials. The hydrocarbon compounds in oil sludge contain benzene, toluene, ethylbenzene, xylene and heavy metals that are potentially carcinogenic [1].

According to the Indonesian government regulation No. 101/2014, oil sludge is classified into the hazardous and toxic waste (B3) with accession number B 351-3. In the present work, oil sludge from PT. Pertamina Balongan is used. The latest laboratory analysis in 2016 revealed that the TPH content in the oil sludge from Balongan was 29.50%, with C/N ratio as much as 149.5. On the other hand, the decree of Indonesian Minister of Environment No.128/ 2003 clearly states that the oil sludge with TPH contents with more than 15% was obliged to be processed prior to its utilization. Nevertheless, Chevron Indonesia argued that the petroleum-contaminated soil could be categorized safe and removed from the treatment site to the environment when the TPH content is less than $\leq 1\%$ [2-4].

Bioremediation could be defined as a natural-clean up process to remove the hazardous chemicals by the microorganisms. During the remediation process, the microbes will degrade the hazardous material and as consequence, harmful products such CO₂ will be produced [5]. Recent studies showed that utilization of various bio-fertilizers could be applied as bio-stimulators to enhance the microbial growth in degrading the oil sludge. Indeed, this concept has been applied in bioremediation process and known as "Bio-fertilizer for Bioremediation". Indeed, Pranajaya [6] reported that compost and biological fertilizer are natural materials and provide higher efficiency in TPH biodegradation as compared to synthetic materials. Moreover, recent work from Juliani and Fudhola [7] demonstrated that compost has a positive effect on the bioremediation process. Briefly, reactor with 10% of compost had significantly higher TPH reduction rate (39%) than that of given only 5%, with the best proportion of mixture between soil and oil sludge as much as 1:1.

In the present work, biostimulation technique will be applied by adding several types of bio-fertilizers, namely the spent mushroom substrate (SMS), vermicompost (VC), domestic compost (DC) and leachate (L), as the source of inoculum and bio-stimulators for the microbial growth to degrade the oil sludge. Additionally, the present study will test different types of bio-stimulators in order to obtain the best inoculum in degrading the Total Petroleum Hydrocarbon (TPH), Polycyclic Aromatic Hydrocarbon (PAH), C/N ratio and heavy metals in oil sludge.

Materials and Methods

Experimental set up

The present work was conducted from February to July 2017 in the laboratory of microbiology, Biology Department, Faculty of Mathematics and Natural Sciences, Padjadjaran University. The analysis of heavy metal content was done at Central Laboratory of Geological Survey, Bandung. While the TPH analysis was conducted at Nutrition Laboratory, Faculty of Animal Husbandry, Universitas Padjadjaran. Additionally, PAH analysis was realized in the Instrument Laboratory, Chemistry Department of Universitas Pendidikan Indonesia, whereas analysis of C/N ratio was conducted in the Laboratory of Soil Chemistry, Faculty of Agriculture Universitas Padjadjaran.

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In this study, experimental method with single factor consisting four levels (the addition of SMS, VC, DC and L to degrade 30% oil sludge) was applied. The main parameters measured in this study were the number of microbial colonies (CFU mL⁻¹), Total Petroleum Hydrocarbon (TPH), Polycyclic Aromatic Hydrocarbon (PAH), and heavy metals content, as well as C/N ratio. TPH and C/N ratio were analyzed within 15 days of interval while PAH and heavy metals were measured at the beginning (day 0) and end (day 45) of experiment during 45 days. On the other hand, supporting parameters consist acidity (pH), temperature (°C) and medium humidity (%).

The measurement of TPH content (%) was performed using the soxhlet method from AOAC in 1995, while C/N ratio (%) was measured according to Walkley and Black' protocol in 1934. Additionally, number of microbial colony (CFU mL⁻¹) was measured using Total Plate Count (TPC) method as described in Cappuccino and Sherman in 1987. Whereas heavy metals content such as Pb, Ni, Cd, Cr and Hg were analyzed using Atomic Absorption Spectrophotometer method (AAS). The PAH content in this study was measured using Gas Chromatography/Mass Spectrometry (GC/MS). Both main and supporting parameters were analyzed descriptively.

Addition of oil sludge medium with inoculum

Medium oil sludge 30% mixed with sand, soil (2:1). In the present study, 25% of each VC, SMS, DC and L were added from their total amount (500 g). The combination applied on each treatments are described in Table 1.

Results and Discussion

Effect of different types of inoculum on TPH content

In the present work, TPH content in the medium containing 30% of oil sludge with the addition of four different types of inoculums were analyzed using Soxhlet during 45 days of experiment. The initial TPH content in 100% of oil sludge was 29.50%. Significant decrease in TPH content occurred in all types of inoculum (Figure 1). The highest diminution in TPH content was found in SMS inoculum (27.7%). On the other hand, DC, VC and L inoculum showed lower percentage in TPH reduction, as much as 24.8%, 20.9% and 24.20%, respectively.

Our results indicate that the SMS inoculum gave significant result in reducing TPH content. According to Black and Hannah mycelium

Treatment	Types of inoculum						
	Oil sludge (%)	Soil (%)	Sand (%)	VC (%)	SMS (%)	DC (%)	L (%)
P1	30	30	15	25	-	-	-
P2	30	30	15	-	25	-	-
P3	30	30	15	-	-	25	-
P4	30	30	15	-	-	-	25

Table 1: Composition of treatment.

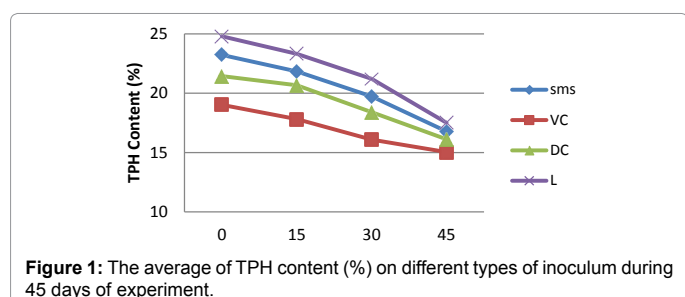


Figure 1: The average of TPH content (%) on different types of inoculum during 45 days of experiment.

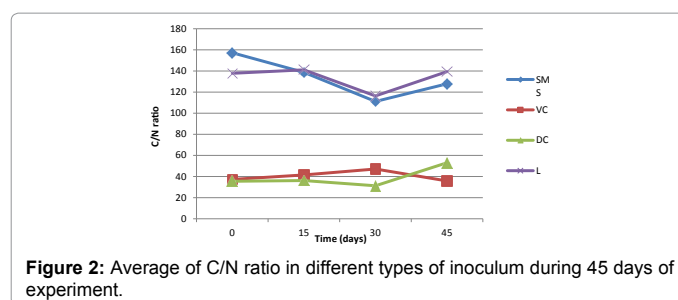


Figure 2: Average of C/N ratio in different types of inoculum during 45 days of experiment.

of white oyster mushroom (*Pleurotus ostreatus*) produces ligninolytic enzymes, including several enzymes that collaborate to decrease lignin. The main enzymes involved in ligninolytic system are lignin peroxidase, laccase, peroxidase and manganese peroxidase. In this case, *P. ostreatus* produce laccase, manganese peroxidase and lignin peroxidase. Furthermore, these ligninolytic enzymes have the ability to break down the complex molecules in crude oil. Therefore, the TPH reduction in SMS inoculum was probably due to the presence of these enzymes. Additionally, difference responses of each inoculum to TPH content during the remediation process could also be due to the variation in nutrient that present in inoculum (Table 1).

According to the criteria from Hardjowigeno [8], Total Nitrogen (N), Phosphorus (P) and Potassium (K) for more than 0.75, 0.035 and 0.006, respectively, are extremely high. These high nutrient content will affect the microbial degradation of hydrocarbons due to their nutritional needs throughout the remediation.

Nutrition is an essential factor in the synthesis dan growth of cells in enzyme activities, which produced by bacteria in degrading pollutants. Indeed, the C, N and P are the essential nutrition for the microorganisms. For instance, Carbon will be used as the source of energy for supporting their activities. While Nitrogen and Phosphorus are the constituents of significant compounds in cells that determine the growth activity of microorganisms. Additionally, the N also has a significant role in the preparation of nucleic acids, amino acids and enzymes. On the other hand, the P element has crucial role in nucleic acid and phospholipid formation. Therefore, these three elements should be presence in appropriate ratio to obtain optimum microbial growth [9].

The microorganisms have the ability to degrade hydrocarbon compounds due to the ezymes that they produce, which enable them to break down the complex organic compounds into simpler compounds. Indeed, enzymes such monooxygenase and dioxygenase are able to open the carbon bonds and produce primary alcohols [10].

Effect of different types of inoculum on C/N ratio

In the present study, the initial C/N ratio in oil sludge 100% was 149.45. Curtailment in C/N ratio during 45 days of remediation is presented in Figure 2. The C/N ratio decreased in SMS and VC inoculum with 18% and 4% of reduction compared to the initial. On the other hand, augmentation in C/N ratio occurred in DC and L inoculum as much as 33% and 1.14%, respectively. The initial C/N ratio reduced in oil sludge medium with VC inoculum after the treatment. Interestingly, this value increased from 15 to 30 days and then decreased to 35.85 in day 45. In contrast, oil sludge with DC and L inoculum showed fluctuation in C/N ratio during the experiment and finally increased at the end of experiment (d 45). Additionally, C/N ratio in SMS inoculum initially increased to 157.24 but then slightly decreased on day 15 and became 127.8 at the end of the experiment (Figure 2).

Our results revealed that DC had higher C/N ratio compared to VC inoculum (16.24 and 15.76, respectively). However, these C/N ratio value in both types of inoculum are still in the allowed level in national standard (SNI 19-7030-2004), which were between 10 and 20. On the other hand, the C/N ratio in L was much lower than that of SNI. In contrast, C/N ratio in SMS medium was much higher that the allowed amount in SNI.

Types of inoculum	Types of nutrient		
	Nitrogen (%)	Phosphorus (%)	Potassium (%)
SMS	0.39	0.24	0.08
VC	2.26	5.03	0.95
DC	2.23	5.10	0.51
L	0.32	0.10	0.87

Table 2: Nutrient content in each types of inoculum.

Types of Inoculum	Parameters measured		
	Total C (%)	Total Nitrogen (%)	C/N ratio
SMS	34.43	0.39	88.28
VC	2.26	2.26	15.76
DC	36.23	2.23	16.24
L	1.59	0.32	4.96

Table 3: Total C, N and C/N ratio at different types of inoculum.

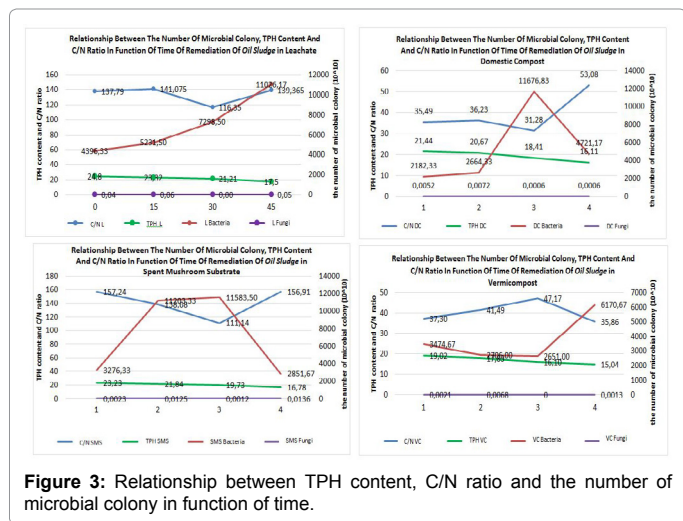


Figure 3: Relationship between TPH content, C/N ratio and the number of microbial colony in function of time.

High value in C/N ratio in SMS inoculum (88.28) causes higher C/N ratio produced during remediation, which is higher than that of VC and DC inoculum. The SMS inoculum contains sawdust, which is a bran as fungal hyphae medium consisting cellulose. According to Ghassani [11], the cell wall of fungi is composed by lignin, which is a carbohydrate-containing polysaccharide with a complex wall structure chain and chitin (C₈H₁₃O₅N)_n. This may explain the high carbon content and C/N ratio observed in the medium (Tables 2 and 3).

In general, decomposition releases carbon dioxide, thus higher microbial activity can accelerate decomposition of organic materials. Therefore, the organic carbon could decrease while the total N increase, which may reduce the C/N ratio. Additionally, the higher total-N formed could also decrease C/N ratio. Indeed, a low C/N ratio indicates that mineralization works well [12].

The degraded polyhydrocarbon will produce new compounds either in a polyaromatic form (ring-shaped hydrocarbon) or a polyalicyclic (double hydrocarbon chain), which may increase or decrease the carbon element. In addition, the N content in oil sludge and all three inoculum cause modification in N element by the microbes hence changing C/N ratio [13].

Correlation between TPH, C/N ratio and the number of microbial colonies in function of remediation time

In this present work, the calculation of the number of microbial colonies (fungi and bacteria) was performed by Total Plate Count (TPC) method. According to Vasudevan and Rajaram [14], the number of microbial population plays crucial role in increasing the rate of hydrocarbon degradation. Indeed, the number of microbial population is positively correlated with the hydrocarbon degradation. In this study, relationship between the number of microbial colony, TPH content and C/N ratio in function of time of remediation of oil sludge 30% is presented in Figure 3.

In remediation of oil sludge 30% with different types of inoculum, the role of bacteria appeared to be more dominant than fungi, indicating that indigenous bacteria in each inoculum were more adaptable in the medium (Figure 3). In each inoculum, the fungal colonies grew actively and increased only on day-0 until day-15, and then the growth was very limited (day-30). On day-45, fungal colony re-grew but still lower than in day-30. The fluctuation in the number of microbial colony in the medium of oil sludge 30% with different types of inoculum indicates that

No	PAH Compounds	C & H	Area/Height (A/H) ke-0 (mm ²)				Area/Height (A/H) ke-45 (mm ²)			
			SMS	VC	DC	L	SMS	VC	DC	L
1	Hexane	C ₆ H ₁₄	1.19	1.21	1.19	1.11	1.14	0.94	1.12	1.13
2	Cyclopentane	C ₅ H ₁₀	1	0.93	1.02	0.99	0.95	TD	0.93	0.98
3	Tetracosane	C ₃₀ H ₆₂	-	2.26	2.23	-	-	TD	TD	-
4	Hexadecane	C ₂₀ H ₄₂	2.29	TD	2.23	2.22	2.12	2.18	2.22	2.19
5	Hexatriacontane	C ₃₆ H ₇₄	2.03	2.38	1.79	2.07	1.94	3.8	2.09	1.88
6	Tetrapentacosan	C ₃₄ H ₇₀	2.6	-	2.29	2.3	1.44	-	2.14	1.72
7	Dotriacontane	C ₃₂ H ₆₆	1.87	-	1.8	3.1	1.49	-	1.98	1.62
8	Pentacosane	C ₂₅ H ₅₂	TD	-	-	-	1.69	-	-	-
9	Pentadecane	C ₁₅ H ₃₂	1.91	TD	-	-	TD	2.18	-	-
10	Tricosane	C ₂₇ H ₅₄	2.11	-	-	2.4	TD	-	-	2.46
11	9-Octadecenoic acid, methyl ester	C ₁₉ H ₃₆ O ₂	-	-	-	1.92	-	-	-	TD
12	Hexacosane	C ₂₆ H ₅₄	-	-	-	1.8	-	-	-	TD
13	Di-9Octadecenoyl)-Glycerol	C ₃₉ H ₇₂ O ₅	-	1.13	-	-	-	2.41	-	-
14	n-Eicosanol	C ₂₀ H ₄₂	-	3.58	-	-	-	2.31	-	-
15	Heptadecane	C ₁₇ H ₃₆	-	TD	-	-	-	1.91	-	-

Table 4: The identified polycyclic aromatic hydrocarbon (PAH) compounds in the medium.

Inoculum	Heavy metals									
	Pb (ppm)		Ni (ppm)		Cd (ppm)		Cr (ppm)		Hg (ppm)	
	H-0	H-45	H-0	H-45	H-0	H-45	H-0	H-45	H-0	H-45
Vermi-compost	24	26	<5	<5	<0.02	<0.02	13	15	92	180
Spent Mushroom substrate	26	75	<5	<5	<0.02	<0.02	18	15	136	132
Domestic compost	22	24	<5	<5	<0.02	<0.02	11	18	146	141
Leachate	34	38	<5	<5	<0.02	<0.02	17	16	170	180

Table 5: The average of heavy metals content at different types of inoculum.

the indigenous microbes in each inoculum were able to adapt and use the hydrocarbon as the source of energy. These microorganisms also utilize the nutrients present in each inoculum for their survival (Table 4).

Fungi are the main component in biodegradation of hydrocarbon. Furthermore, biodegradation product by the fungal organism will be utilized by the bacteria thus facilitating the competition of growth between the fungi and bacteria in the medium. Indeed, the high population of bacteria suppresses the growth of fungi that may decrease the fungal population [13].

The TPH content decreased during the remediation. This reduction may lead to the formation of new toxic compounds such as polyaromatic hydrocarbon (PAH) thus lowering the number of microbial colony and provokes the augmentation in C/N ratio. According to Rossiana [13], the degraded polyhydrocarbon could produce new compounds in both polyaromatic form and polyalicyclic that probably may increase or decrease the carbon content.

Effect of different types of inoculum on polycyclic aromatic hydrocarbon

The PAH content was analyzed at the beginning (day 0) and the end of remediation experiment (day 45). In this study, a total of 15 hydrocarbon compounds were detected in all four types of inoculum that consist of; Hexane, Cyclopentane, Hexadecane, Hexatriacontane, Dotriacontane, Tetrapentacosan and Dotriacontane. These compounds have a range of carbon chains from C₆-C₃₄. Indeed, change in proportion of each of these compounds during remediation was occurred (Table 5).

Degradation of Pentadecane and Tricosane compounds was detected in SMS medium (Table 5), until both of them were undetected at day 45 and replaced by new compound (Pentacosane). On the other hand, the Cyclopentane and Tetracosane compounds were undetected in VC inoculum, but three other new compounds (Hexadecane, Pentadecane and Heptadecane) appeared at day 45. In DC inoculum, only one compound (Tetracosane) that was undetected at day 45, while in L inoculum the 9-Octadecenoic acid, methyl ester and Hexacosane remain undetected. Furthermore, other compounds found in oil sludge medium decreased qualitatively based on the observation of their surface area from the beginning until the end of observation.

In this study, the change in the range of carbon chains and the amount of detected compounds during remediation indicate degradation of complex carbon chains into the non-complex carbon chains as consequence of microbial activity. This result confirms the previous study by Nugroho [10] that the presence of microbial activity could degrade oil sludge by fragmenting the long-chain aliphatic hydrocarbon component and transforming the aromatic hydrocarbon compound. Therefore, change in composition of the constituent hydrocarbon fraction can be revealed in oil sludge. Indeed, fragmentation in the heavy fraction of hydrocarbon bonds may affect a rapid multiplication in lighter fraction of hydrocarbon.

Degradation of PAH is closely related to the mechanism of action of dioxygenase and dehydrogenase enzymes. The first step in PAH degradation involves oxygen atoms on 2 carbon in the structure of benzene by dioxygenase enzyme. This interaction produces the structure of cis-dihydriol. Afterwards, dehydrogenase enzyme can trigger re-aromatization on the structure, thus the dihydroxylated intermediat is formed. This structure then undergo the cleavage of aromatic ring to form the tricarboxylic acid cycle (TCA) intermediat, which subsequently become non-complex structure thus it can be processed into the source of energy [15].

Effect of different types of inoculum on heavy metals

In this study, heavy metals were analyzed at the beginning (day 0) and end of the remediation experiment (day 45). According to the Government Regulation of Indonesian Republic No. 85/1999 on the management of hazardous and toxic wastes, the maximum allowed limit of Pb, Ni, Cd, Cr and Hg were 5.0 ppm, 0.4 ppm, 1.0 ppm, 5.0 ppm and 0.2 ppm, respectively. However, our study revealed that most of these heavy metals (except Cd) exceeded the allowed limit.

In general, Pb content in all types of inoculum in oil sludge medium increased. Indeed, the highest augmentation in Pb content was found on SMS inoculum (188%) as compared to other types of inoculum. The increased in Pb content for VC, DC and L were as much as 8.33%, 9.09% and 11.7%, respectively. In contrast, there was no change in both Ni and Cd content from the beginning (day 0) to the end of experiment (day 45). The content of Cr decreased in both oil sludge medium with SMS and L inoculum as much as 16% and 5.88%, respectively whereas the increase of Cr was observed in both VC (15%) and DC (63.6%) inoculum. Interestingly, the Hg content increased only in SMS inoculum (29%), but not on DC, L and VC inoculum.

Reduction in heavy metals content occurs naturally due to the presence of microorganisms that are able to grow, adapt and reproduce in their environment. Indeed, the dominant microorganisms that developed during remediation are the microbes. According to Vijayaraghavan and Sang Yun, bacteria are considered as heavy metals biosorbent agent. Indeed, the mechanism of heavy metal biosorption by microorganisms is a complex process that consists of transportation of heavy metals through cell membranes, ion exchange and production of organic acids by the bacteria. Additionally, Sayer and Gadd observe that one of chelator produced by fungi is oxalic acid. This type of acid is produced by microbes and can increase microbial resistance to metals via the formation of insoluble metal-oxalic complexes.

Evelyn [16] observed that the enlargement of heavy metals content in the medium is perhaps due to the accumulation of dead metals by the bacteria then these metals were released back to bind with the soil colloids thus increasing the metal content in the soil.

Effect of acidity (pH), temperature and humidity of medium in each inoculum

The main component in bioremediation is the bacteria as biodegradation agent and the product produced such as converting or mineralizing the contaminants thereby reducing their mass and toxicity that are different from most other components on their environment. The success of bioremediation is influenced by three crucial factors: the microbial availability, contaminants accessibility and favorable environment. Indeed, the efficiency of bioremediation depends on the ability of microbes to degrade complex compounds to non-complex compounds [17].

The supporting parameters measured in this study are temperature, acidity and humidity. The temperature in all treatments during remediation ranged from 25°C to 28°C. Udiharto reported that the optimum temperature for bioremediation of oil sludge range from 25°C to 40°C, due to the requirement of mesophilic bacteria that live in that range of temperature in order to enhance bioremediation. Thus, the temperature obtained in this study are still in optimum range. On the other hand, our results revealed that the pH in the medium decreased after 6 weeks of remediation. This diminution is probably due to the bacterial metabolism activity in oil degradation that produce fatty acids as final products [18].

In the present work, the average of humidity ranged from 47% to 54%. According to Cookson [19], the optimum humidity for bioremediation of oil sludge and its derivatives is approximately 50% to 60%. Moreover, inadequate humidity (i.e., less than 40%) could limit the rate of bioremediation, while the soil humidity above 70% could disrupt oxygen transfer thus reducing aerobic activity.

Conclusion

Our results revealed that the four types of inoculum: SMS, VC, DC and L, affect the decrease of TPH content as much as 27.7%, 20.9%, 24.8% and 24.20%, respectively. The SMS and VC are able to reduce C/N ratio as much as 18 and 4%, respectively. While the DC and L inoculum increased C/N ratio as much as 33% and 1.14%, respectively. The SMS and L inoculum decrease the Cr content as much as 16% and 5.88%, respectively whereas Hg content reduced to 29% in oil sludge with SMS inoculum. Furthermore, PAH could be degraded by all types of inoculum, from the complex carbon chain (C₆-C₅₄) into the non-complex carbon chain (C₆-C₃₀). Our results suggest that longer period of time required degrading oil sludge 30% to significantly reduce TPH, C/N ratio, PAH and heavy metals content as regulated in the national standard. In addition, further studies are required to identify the microbes in each inoculum that affect biodegradation of oil sludge.

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